

OPTIMIZATION OF BIOGAS PRODUCTION FROM ORGANIC FRACTION OF MUNICIPAL SOLID WASTE: EXPERIMENTAL TEST USING LIQUID FROM WASTE FERMENTATION

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Abstract: The present study aimed to produce biogas using liquid from OFMSW fermentation. In this study the physic-chemical parameters of liquid from OFMSW were determined. Anaerobic digestion using liquid from OFMSW and 2 % (w/v) of OFMSW as substrate load was carried out both in batch anaerobic digesters during 25 days. The pH, ammonium nitrogen and gas (CO₂, CH₄ and H₂S) were followed during anaerobic digestion. The concentration of ammonium nitrogen ranged respectively 3 g NH₄-N·L⁻¹ to 6 g NH₄-N·L⁻¹ and 5 g NH₄-N·L⁻¹ to 9 g NH₄-N·L⁻¹ in anaerobic digestion from liquid from OFMSW and 2 % of OFMSW. This technology has permit to obtain good quality of biogas (57 % CH₄, 3 % CO₂ and 20 ppm H₂S) in a short time of 20 days.

Keywords: anaerobic digestion, biogas, liquid from OFMSW, Municipal Solid Waste

INTRODUCTION

The world knows today an increase in municipal solid waste (MSW) generation with ever-increasing populations. According to Amoo and Fagbenle [1], as of 2011 the world generates estimated 2 billion tons of municipal solid wastes each year. Almost 54 % of the US's annual production of about 243 million tons [2] is still landfilled. Similarly in 2008, Canadians produced 34 million tons of MSW of which 76 % was disposed in landfills [3]. In Burkina Faso, precisely in Ouagadougou solid waste production is estimated at 300 000 tons per year since 2012 by sustainable development department. Only 50 % of waste is collected and disposed directly to the landfill. None evacuated waste pile up on vacant lots or in uncontrolled landfills. As a direct result of environmental pollution and degradation of living conditions (including health) populations (microbial contamination, infestation of the atmospheric air by noxious odors, etc.). This form of management is not sustainable and leads to production of leachate and uncontrolled greenhouse gas emissions. So, energy issue in Burkina Faso is of great concern and hinders its development. The government has taken steps to diversify the energy mix and put the emphasis on renewable. The organic fraction of municipal solid waste (OFMSW) is a large and renewable potential energy source that can be exploited on a sustained basis if treated under controlled conditions to reduce the environmental impact and recover energy; this is demonstrated by many studies [4 – 7]. Anaerobic digestion is a most cost-effective bioconversion technology that has been implemented worldwide for commercial production of electricity, heat, and compressed natural gas from organic materials [8, 9]. Anaerobic digestion is a biological process wherein diverse group of microorganism convert the complex organic matter into a simple and stable end products in the absence of oxygen [10].

This processes participle strongly on the other hand in the management of waste for bioconversion of different high-strength biowaste, such as municipal primary and secondary sludge, organic portion of kitchen waste, pulp and paper sludge, agricultural off farm and on-farm residues including animal manure into methane-rich biogas [11 – 13]. Anaerobic digestion is all biochemical reactions that involve several phases [14, 15]. The first phase, hydrolysis and acidogenic are ensured by hydrolytic bacteria. These bacteria convert the complex organic matter (lipids, cellulose, starch, protein...) into simpler compounds, volatile fatty acids (AVG) (acetic acid, propionic acid, butyric acid ...) and alcohols (methanol, ethanol ...). In this phase there is also the production of hydrogen (H_2) and carbon dioxide (CO_2) from the reduction of lipids and proteins. During the second phase (acidogenesis), the products are converted into acetate (CH_3COO^-) and hydrogen (H_2). During the third phase, methanogenic bacteria hydrogenophiles reduce CO_2 to methane (CH_4) using hydrogen and acetoclastes methanogenic bacteria convert acetate to CH_4 . These phases are a series of interlinked reactions proceeding spatially as well as temporally in consecutive and parallel steps and hence, influence one another [16]. According to several author, hydrolysis is the limiting step of anaerobic digestion process [14, 15]. When methanogenesis is not rapid enough due to some upset, volatile fatty acids (VFA) accumulate, which may lead to a decrease in pH and a cessation of the methane production. Returning to normal operating conditions can be costly a time-consuming. Remedy actions include reducing the organic loading rate to the point where the VFA production rate is less than their maximum consumption rate. This which allow for consumption of the excesses VFA

and a return of neutral pH until stable conditions occur again. If this measure is not sufficient, decrease in loading must be coupled with the addition of appropriate chemicals for pH correction [17, 18].

Many optimization techniques have been developed in recent years to minimize acidification during the methanization and increase biogas production. It's the case of alkali pretreatment process prior to anaerobic digestion [19], recirculation leachate in batch mode solid-state anaerobic digestion [20]. Other studies realize separation of hydrolysis phase; it is a pretreatment before anaerobic digestion step [9 – 21]. Physical and chemical pretreatments of waste are expensive because demand a lot of energy. Also according to Bisaria and Ghose [22] thermo-chemical treatment or physico-chemical properties of plant polymers generate hydroxymethylfurfural and furfural compounds which are highly toxic to microorganisms.

Biological pretreatment includes both anaerobic and aerobic methods, as well as the addition of specific enzymes such as peptidase, carbohydrase and lipase to the anaerobic digestion system [23]. Carrère *et al.* [24] and Ge *et al.* [25] considered the hydrolytic-acidogenic step (first step) of a two-phase anaerobic digestion process as a biological pretreatment method. This treatment can result in a higher methane production [26]. Others studies consider it as a process configuration of anaerobic digestion, but not a pretreatment method [27, 28].

The objective of this study is to provide a waste by anaerobic digestion processing technology, using liquid from OFMSW fermentation.

MATERIALS AND METHODS

Sampling and preparation for experiment

Sampling of municipal solid waste was carried out in three (3) centers pre-collects of Ouagadougou's town. Each tank three trash (3) repetitions were performed. The first center is located in the district 2 (North latitude: 12° 22'; West longitude: 1° 32' and altitude: 335 m), the 2nd in the district 12 (North latitude: 12° 19', West longitude: 1° 31' and altitude: 349 m), and the third center in the district 3 (North latitude: 12° 23'; west longitude: 1° 32' and altitude: 326 m). Waste sampled were mixed, sorted and dried in the sun for 7 days, then crushed and sieved (size ≤ 1 mm).

Determination of physical chemical parameters of wastes

The pH , acidity, conductivity and salinity were measured using a WTW's digital multi-parameter system (inoLab Multi 9420 IDS, Germany). Dry matter (DM) was determined according to standardized method AFNOR NF U 44-171 of October 1982. 5 g of quantity of sample was placed in the oven (POL-EKO-APARATURA, Pologne) at 105 ± 2 °C until constant weight, approximately 24 hours. Mineral matter or total was determined using TAPPI T 211 CM-86 norm [29]. 1 g of waste powder was introduced into a crucible (Nabertherm GmbH, Germany) and placed in the oven at 575 °C for 3 h. The organic matter content in a sample was obtained by the difference between dry matter and total ash (mineral matter). Klason method according to Tappi Standard 222 [30] was used to determine lignin tenor. Lignin tenor was performed with

0.1 g of OFMSW in 1.5 mL of 72 % H_2SO_4 in a 50 mL polypropylene centrifuge tube and let to act for 2 h at 20 °C. The mixture was diluted with 56.2 mL of distilled water and filtered under vacuum using the previously weighed crucible, washing with distilled water and drying in a ventilated oven. Calorific value of municipal solid organic waste sample was determined using the e2k bomb calorimeter systems combustion calorimeter manufactured by Digital Data Systems (Pty) Ltd (Brownstone Asia) device.

Anaerobic digestion

Inoculums preparation and liquid from OFMSW

Cow dung was used as inoculums after enrichment. The medium component was NH_4Cl 2 g·L⁻¹, K_2HPO_4 2 g·L⁻¹, NaCl 13 g·L⁻¹, CaCl_2 6H₂O 0.17 g·L⁻¹, FeCl_3 0.1 g·L⁻¹, CuSO_4 5H₂O 0.1 g·L⁻¹, acetic acid 1.5 g·L⁻¹, sodium acetate 5 g·L⁻¹, peptone 1 g·L⁻¹. The pH was adjusted at 7 ± 0.2 with NaOH 2N. Enrichment was performed during 6 days.

The prefermentation step of waste powder was conducted in a 20 L vessel for 10 days. The substrate concentration (waste powder) was 15 % (w/v) in a reaction medium of composition: 2 g K_2HPO_4 and 2 g NH_4Cl to 1000 mL. At the end of fermentation, the reaction phase was filtered in order to obtain the liquid.

Anaerobic batch test

Batch test were conducted in glass bottles as anaerobic digester (capacity 1000 mL). The experimental design used Angelidaki method [31] and was showed in Figure 1. In digester N°1, 500 mL of liquid from OFMSW was used as substrate with inoculums 10% (v/v). In digester N°2 (2 % (m/v) of OFMSW), OFMSW was taken as 10 g in 500 mL of phosphate buffer (2 g K_2HPO_4 and 2 g NH_4Cl in 1000 mL distilled water) and 50 mL of inoculums (10 % (v/v)). Digester N°3 was used as control (without substrate). It was composed only by 500 mL phosphate buffer and 10 % (v/v) of inoculums.

The biogas volume from the reactors was collected in bag and quantified. Digesters were incubated at 37 °C.

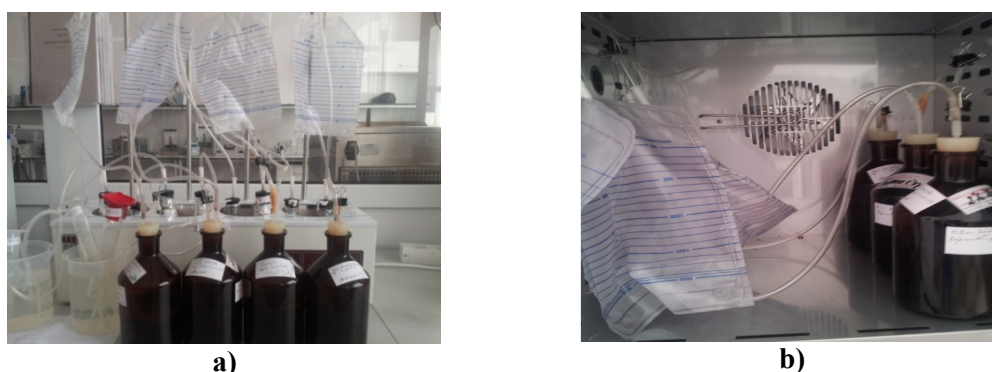


Figure 1. Pictures of reactors setup: (a) view; (b) incubation at 37 °C

Chemicals products were provided by S.C. Chim Reactiv, S.R.L (Romania).

Monitoring anaerobic digestion

The ammonia content ($\text{NH}_4\text{-N}$) was analyzed by Spectroquant® photometer NOVA 60 from Merck Millipore (United States) following the instrument's protocol. The gas composition was analyzed using Biogas 5000 from Geotech (France).

RESULTS AND DISCUSSION

Physico-chemical characteristics of OFMSW sample

Characteristic of OFMSW is shown in Table 1. Characteristic present 92.79 % of Total solid, 80.10 % of Total volatile solid, 29.21 % of lignin, 17.11 % of ash and C/N rapport 31.65. The very high Total solid is due to pretreatment of waste, sorting, drying and reduced sizes. Total volatile solid (80.10 %) was values similar to those previously reported [32, 33].

The high proportion of lignin in the sample (29.21 %) is due to presence of rich waste that matter like grass debris, leaves and branches. According to Eleazer *et al.* [34], these components of the waste have high lignin 5 to 43.8 %. Lignocellulosic biomass has complexity and variability chemical structures; this substrate is resist to biodegradation. The off for outers as Hendriks and Zeeman [35], Zheng *et al.* [9] report necessity to pretreated recalcitrant lignocellulosic biomass before anaerobic digestion process. According to Zhou *et al.* [21] our sample can be classified in food residue samples, wood waste samples and rubber. In this group, ash content varies from 7 % to 19 %; volatile mater varies from 63 % to 74 %; the fixed carbon content varies from 15 % to 26 % and heating value locates between 20 and 25 $\text{MJ}\cdot\text{kg}^{-1}$. The C/N ratio of the sample 31.65 is favorable for biodegradation because it is very proximate to the optimum between 25 and 30 [36, 37]. These chemical profiles make samples good candidates for methanization.

Table 1. Physico-chemical characteristics of OFMSW

Composition	Average
Total solid (TS) [%]	92.79 ± 0.09
Total volatile solid [% TS]	80.10 ± 0.32
Total ash [% TS]	17.11 ± 0.30
Combustible [% TS]	46.57 ± 0.32
Lignin [% TS]	29.21 ± 0.03
CV [$\text{MJ}\cdot\text{kg}^{-1}$]	17.48 – 11.49
C/N	31.65 ± 0.05

Physico-chemical characteristics of extract juice

Table 2 shows the physico-chemical characteristics of liquid from OFMSW was compared to organic fraction of MSW, waste dairy and Juice apples fermented used for anaerobic digestion. Liquid extract from OFMSW present respectively 5.21, 4.90 $\text{g}\cdot\text{L}^{-1}$, 8.70 $\text{mS}\cdot\text{cm}^{-1}$, 8.69 $\text{g}\cdot\text{L}^{-1}$ and 5.36 $\text{g}\cdot\text{L}^{-1}$ for pH, salinity, conductivity, Total Dissolved Solutes (TDS) and ammonium concentration ($\text{NH}_4\text{-N}$). The main of these parameters were higher than value found by El-Chakhtoura *et al.* [38] and Quezada [39] for

OFMSW, waste dairy and apples fermented juice. Moletta [13] conducted tests on anaerobic digestion of marine microalgae, obtained highest yield of CH_4 with $15 \text{ g}\cdot\text{L}^{-1}$ of salinity. The increase of salinity is implied in decrease of CH_4 yield. Chen *et al.* [40] showed that salinity $15 \text{ g}\cdot\text{L}^{-1}$ corresponding to concentration of $4.6 \text{ g}\cdot\text{L}^{-1}$ of NaCl which was lower to toxic concentration recommended. According to Soto *et al.* [41] sodium was more toxic than propionic acid and acetic acid coming of volatile fatty acid deterioration during bacteria metabolism. The extract juice which constitutes substrate in this study had lower salinity values to $15 \text{ g}\cdot\text{L}^{-1}$ suitable for anaerobic digestion. Zeng *et al.* [42] found similar TDS values $6.80 \text{ g}\cdot\text{L}^{-1}$ but a high conductivity of $13.61 \text{ mS}\cdot\text{cm}^{-1}$ during anaerobic digestion of animal dung. Graterol [43] found conductivity and TDS values respectively of 2.58 at $4.28 \text{ mS}\cdot\text{cm}^{-1}$ and 16.13 at $26.75 \text{ g}\cdot\text{L}^{-1}$ in anaerobic digestion of industrial wastewater.

Table 2. Physico-chemical characteristics of liquid from OFMSW

Parameters	Average
pH	5.21 ± 0.00
Salinity [$\text{g}\cdot\text{L}^{-1}$]	4.90 ± 0.00
Conductivity [$\text{mS}\cdot\text{cm}^{-1}$]	8.70 ± 0.05
Total dissolved solutes [$\text{g}\cdot\text{L}^{-1}$]	8.69 ± 0.04

Monitoring of pH and $\text{NH}_4\text{-N}$ during anaerobic digestion

Figure 2 shows the change in pH over time depending on the type of treatment. The pH was adjusted for all the reactors close to the neutral pH by adding 2N NaOH to the feeding solution during the incubation. Only the digester N°3 (control) did not show a significant change in pH. This can be explained by the absence of substrate in medium. Milaiti *et al.* [44] and Nikiema *et al.* [18] showed weak acidification with 2 % load substrate. Liquid extract from OFMSW has the same characteristic with 2 % load substrate; the pH is stabilized at 10th day of process. Bacteria deteriorate gradually organic macromolecules and free organic acid in the medium.

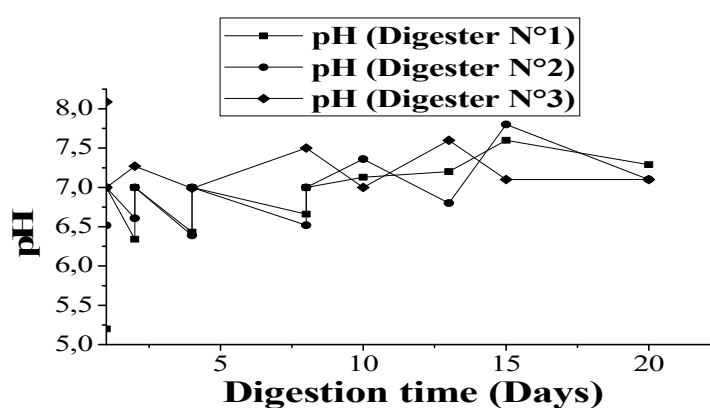


Figure 2. pH variation during anaerobic digestion

Ammonium nitrogen is essential to microbial metabolisms and is used by the bacteria for growth. The excess of ammoniacal nitrogen, not used by the microorganisms remains in the form of a mixture of ammonia (NH_3) and ammonium (NH_4^+) in acid-base

balance according pH -dependent and temperature. An optimal ammonia concentration ensures sufficient buffering capacity in the anaerobic digestion system thus increasing the stability of process, while high ammonia is reported as a strong inhibitor of biogas production [40, 45]. In this study, the concentration of ammonium nitrogen range from respectively $3 \text{ g NH}_4\text{-N}\cdot\text{L}^{-1}$ - $6 \text{ g NH}_4\text{-N}\cdot\text{L}^{-1}$ and $5 \text{ g NH}_4\text{-N}\cdot\text{L}^{-1}$ - $9 \text{ g NH}_4\text{-N}\cdot\text{L}^{-1}$ in digester N°1 and digester N°2 (Figure 3). Chen *et al.* [40] reported that the inhibiting concentration of total ammonia nitrogen (TAN) which decrease methane yield at 50 % was ranged from 1.7 to $14 \text{ g}\cdot\text{L}^{-1}$. Our values were correlated to this range. Some off for have mentioned tan anaerobic digester operated with best yield even concentrations of $\text{NH}_4^+\text{-N}$ ranged from 6 to $8 \text{ g NH}_4^+\text{-N}\cdot\text{L}^{-1}$ [46].

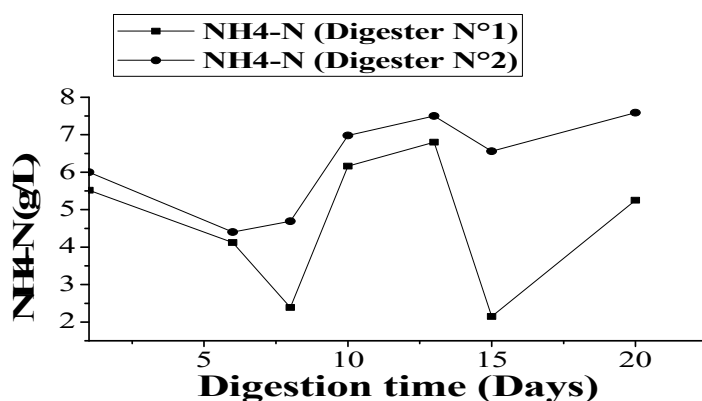


Figure 3. $\text{NH}_4\text{-N}$ variations during anaerobic digestion

Evolution of biogas during anaerobic digestion

The preparations and waste pretreatment aim to change the waste and improve the parameters of degradation, so the quality of biogas and biomethane yield will be positively impacted. The concentrations of hydrogen sulfide of different types of treatment are shown in Figure 4.

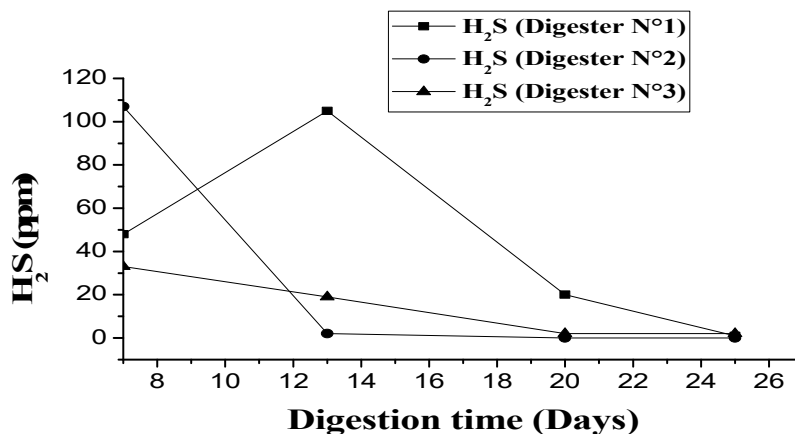


Figure 4. Evolution of H_2S during anaerobic digestion

The aerobic pre-treatment has an impact on the production of H_2S . The H_2S concentration increases from 48 ppm to 120 ppm before decreasing the 20th day to 20 ppm in digester N°1. In digester N°2, the level of H_2S was downed 107 to 2 ppm and completely reduced at a 20th day.

Briand and Morand [47] mentioned that a level of H_2S measured in the biogas produced during anaerobic digestion was about 10.000 – 20.000 ppm. Among the undesirable substances present in the biogas, hydrogen sulfide is the most problematic element because of its toxicity and corrosively action on the equipment. Also its combustion leads to the production of sulfur dioxide [48]. The concentrations of H_2S obtained in our study are below to the maximum concentration of H_2S specified by co-generator manufacturers which is around $150 \text{ mg} \cdot \text{m}^{-3}$ (equivalent to 100 ppm) [49].

The evolution of CH_4 and CO_2 from different types of treatment was showed in Figure 5. According to Burke [50] dairy waste biogas will typically be composed of 55 to 65 % methane and 35 to 45 % carbon dioxide. The monitoring of biogas production was carried out during 25 days. Extract juice gave a maximum of 57 % CH_4 at the 20th day (Figure 5). The fall in production of CH_4 in the case of extract juice is due to substrate depletion in the medium at the 25 days.

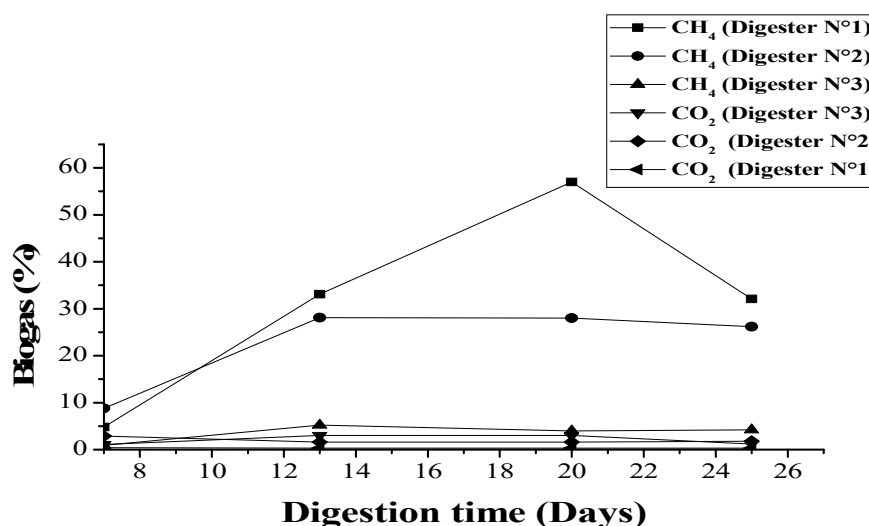


Figure 5. Evolution of biogas (CH_4 , CO_2) during anaerobic digestion

After fermentation, there is a production of organic compounds which are been used during the first 20 days of anaerobic digestion. Parawira *et al.* [51] reported that hydrolysis stage could stimulate the acidogenic microorganisms to produce more specific enzymes, thus resulting in more extended degradation of substrates. At level of 2 % (m/v) of solid waste, production of CH_4 reached 28 % at 20th day and became constant (Figure 5). This can be explained by the fact that the solid substrate in the batch system is continually degraded by microorganisms and used over time. Production of methane in the 20th day in digester N°1, 2 and 3 was showed in Table 3. Analysis Of Variance (ANOVA) at probability level of 0.05 and the Student Newman-Keuls (SNK) tests using the Statistical Analysis System (SAS, Version 8.1) software show very highly significant different ($p < 0.0001$) with CH_4 production in digester N°1 (Liquid from OFMSW), digester N°2 (2 % of waste) and digester N°3 (Control).

Table 3. Methane production in the 20th day

Digesters	CH ₄ [%]	<i>p</i>
Digester N°1	57	< 0.0001
Digester N°2	28	
Digester N°3	5,2	

Another observation made was the low production of CO₂ (1 to 3 %). This low level of CO₂ content in the biogas could be explained by a long period of aerobic prefermentation (10 days). According to Dictor *et al.* [52], at the first 20 days of biogas production, the gas was composed of 5 - 6 % CH₄, 72 % CO₂ and 10 % H₂S. They added that methanogens phase takes place after 36th days. Also CH₄ and CO₂ production was stabilized with proportions respectively between 56 - 75 % and 32 - 42 %. During the 20 days of production using liquid from OFMSW the proportion of CH₄ and CO₂ were range respectively 57 % and 3 %. Wang *et al.* [53] showed that the time from 0 to 20 days is a period of adaptation of microorganisms to the substrate, after that, the stabilization happen at 21 days. Our experiment with liquid from OFMSW fermentation has presented a short time of adaptation.

CONCLUSIONS

The results showed that liquid from OFMSW fermentation can be anaerobically treated well and could be a good potential source for biogas production. The physico-chemical characteristics of substrate showed that it could be compared to animal dung, industrial wastewater, fermented juice apples and dairy waste concerning its utilization for anaerobic digestion. Good quality of biogas (57 % CH₄, 3 % CO₂ and 20 ppm H₂S) has been obtained in a relatively short time of 20 days. It is preferable to perform this type of treatment to avoid overloads that would create shutdown process as a result of acidification. Also this technology could help treat high quantity of municipal organic waste.

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